

University of West Hungary

Theses of PhD dissertation

**Investigation of biocatalytic processes
in the environmental strategy**

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Introduction

Bioremediation technologies have been used widespread in the biotechnological field of environmental applications during the preceding decades, which are based on the degradation of contaminants by enzymes from the microorganisms. The biocatalytic or bioremediation technologies have several advantages over the other physical or chemical ways of rehabilitation, e.g. relative cost-effectiveness; using them *in situ* causes less damage to the original function of land and they can be well connected with other environmental processes, respectively.

However, it should be said that the applicability of these microbial operations are limited by numerous factors, e.g. they are time-consuming and sensitive to environmental conditions. The most important factor from the point of view of my research work is their dependency from peculiarities of the contaminated media and contaminants.

An effective execution and optimization of the biocatalytic environmental strategies requires clear understanding of the biochemical degradation processes.

The aim of my doctoral thesis was therefore to investigate the detection of substrate inhibition and activation effects that are capable of influencing the rate of enzymatic degradation using model and experimental data series.

Calculation- and experimental methods

Modelling of biodegradation

The biodegradation reactions are described by the Michaelis–Menten equation:

$$\frac{d[S]}{dt} = \left(-\frac{d[P]}{dt} \right) = -v_{max} \cdot \frac{[S]}{K_M + [S]} , \quad (1)$$

where S – the concentration of model contaminant, as initial substrate (in mmol/L), P – the concentration of product, K_M – the Michaelis-constant of the biodegradation step, v_{max} – the maximal reaction rate of the biochemical reaction. The results of the model biocatalytic degradation curve are depicted in *Fig. 1*. These kinetic results are involved the feasible mechanisms of various reversible inhibitions and enzyme activation during demolition of the model substrates. Different solutions of degradation model have also been generated by Scilab algorithms. A realistic noise (with different amplitude of concentration mean) was superposed to the data series of the substrates.

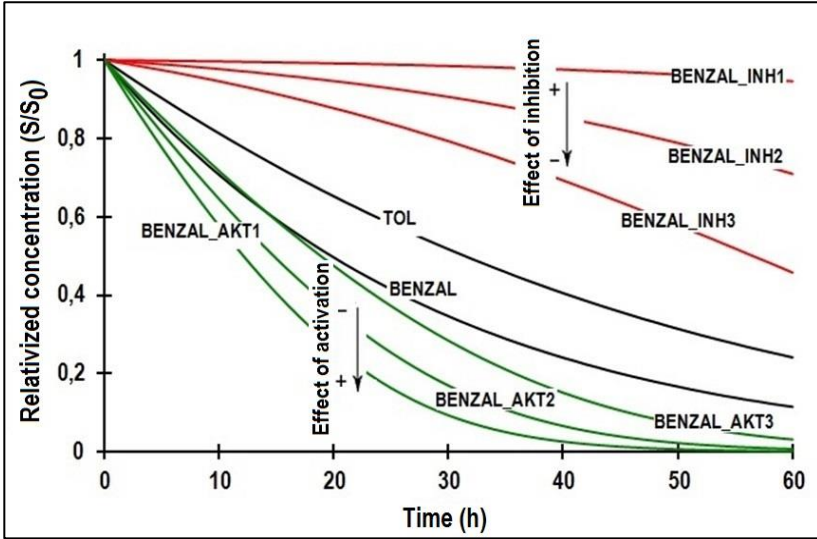


Figure 1. Relativized kinetic curves of model pollutants as a function of time. BENZAL and TOL stand for benzaldehyde and toluene, respectively. The kinetic curves were calculated without inhibition. BENZAL_INH1-2-3 show inhibition effects of varying strength due to the interaction with toluene. BENZAL_AKT1-2-3 show activation effects of varying strength due to the interaction with K^+ ion.

Estimation of the kinetic parameters by regression models

The relation (1) is a first order difference equation, which is transformable at the $t = 0$, $S = S_0$ initial value to the following (1.1) implicit equation:

$$K_M \cdot \ln \frac{[S]}{[S_0]} + S - S_0 = v_{max} \cdot t , \quad (1.1)$$

On the basis of the (1) and (1.1) equation, the values of kinetic parameters (K_M and v_{max}) can be concluded from the measuring/model results of the S substrate. After the Lineweaver–Burk linearization of the (1) equation, the dt/dS and the $1/S$ quantities correlate with each other. The K_M and v_{max} kinetic parameters can be estimated from the regression slope and the intercept of the reciprocal data. In the case of this eval-

uation the numerical derivation of the $S = S(t)$ kinetic curves is necessary. According to the S implicit term of equation (1.1) it is transformable into a quasi-multivariate, linear relation. Using the following variables, $Y = \ln(S/S_0)$ és $X = S$ we can write this formula:

$$K_M \cdot Y + X - S_0 = -v_{max} \cdot t , \quad (1.1.1)$$

which determines a three-dimensional (Y, X, t) line. The multivariate regression coefficients of the (1.1.1) equation provide opportunity to estimate the kinetic parameters by this alternative method. The estimation accuracy of the two different methods was tested in my research using different extents of noise effect and different initial substrate concentration.

Biocatalytic experiments

Potential environmental pollutants, e.g. benzol, toluene, benzaldehyde were applied in the biocatalytic experiments, besides, they were degraded by extracts of earthworm and also by a specific breakdown product. The biodegradation of pollutants was investigated individually as well as in mixtures – in case of inhibition effects – and in the presence of KCl or glutation for the sake of the activation effects. The bioconversions of the pollutants were quantified using gas chromatography.

New scientific results

Thesis 1

The kinetic parameter estimation of the biochemical processes can be executed more precisely using the elaborated multivariate linear regression method in contradiction to the Lineweaver-Burk double reciprocal plot.

The multivariate linear approach (1.1.1 equation) eliminates the distortion errors of derival calculation and its results so that the extreme big outliers cannot manifest in the database. This method has an additional advantage; it can partly decrease the applied noise effect. On the basis of all this the multivariate linear approach enables to execute more precise estimation. This interplay is illustrated by Figures 2.a-2.b in case of the prediction $-(K_M/v_{max})$ kinetic parameter using the two different estimation methods.

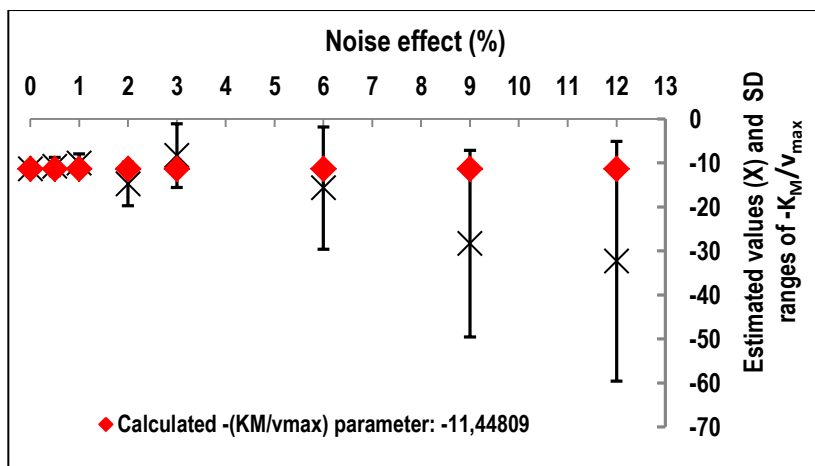


Figure 2.a. Estimation of the $-(K_M/v_{max})$ parameter and their standard deviation ranges using the multivariate linear approach with different noise effect

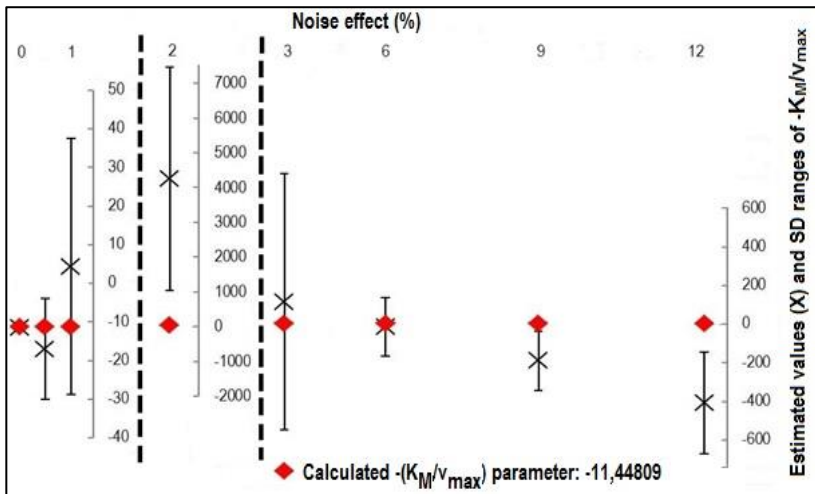


Figure 2.b. Estimation of the $-(K_M/v_{max})$ parameter and their standard deviation ranges using the Lineweaver-Burk approach with different noise effect

Thesis 2

The theoretical error analysis of Δr_v (error of reciprocal maximal reaction rate), namely

$$\Delta r_v = \Delta h_t \cdot \frac{K_M}{S_0} + \left(\frac{K_M}{S \cdot S_0} + \frac{K_M}{S_0^2} + \frac{S}{S_0} \right) \cdot \Delta h_s + \frac{K_M}{S_0} \cdot \ln \left(\frac{S}{S_0} \right) \cdot \frac{\Delta K_v}{K_v}$$

was verified using model data also. In case of lower substrate concentration the precision of the estimation depends on the measuring error of substrate concentration.

The dependence of the maximal reaction rate (v_{\max}) estimation on the initial substrate concentration is depicted in the *Figure 3*.

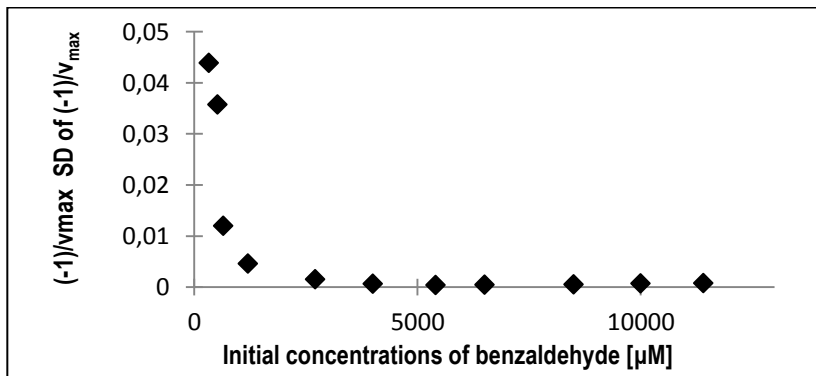


Figure 3. The effect of benzaldehyde initial substrate concentration on the estimation of $-1/v_{\max}$ kinetic parameter standard deviation values. The noise effect is 5%.

This conclusion indicates also one of the weaknesses of the Lineweaver–Burk linearization method, namely that the biggest measuring errors of the initial reciprocal reaction rate will occur at the smallest substrate concentration, thereby the parameters of fitting straight line will be modified and the estimation will become inaccurate.

Thesis 3

Temporal data series of pollutant concentrations that are resulted by similar degradation mechanism (single or mixed) can be linearly correlated to each other. The parameters of fitting regression straight line (slope and intercept) are changed significantly by the enzyme modification effect.

According to the alteration of reference regression straight line parameters the inhibition or activation effects can be determined. In case of inhibition the slope of straight line is increasing and the intercept is decreasing while in case of activation the alteration of the parameters occurs in the opposite direction. The alteration of the regression parameters is a measure of the strength of the modification effect (HERKE et al. 2015). *Figure 4.a and 4.b* illustrate the results in case of model inhibition and experimental data.

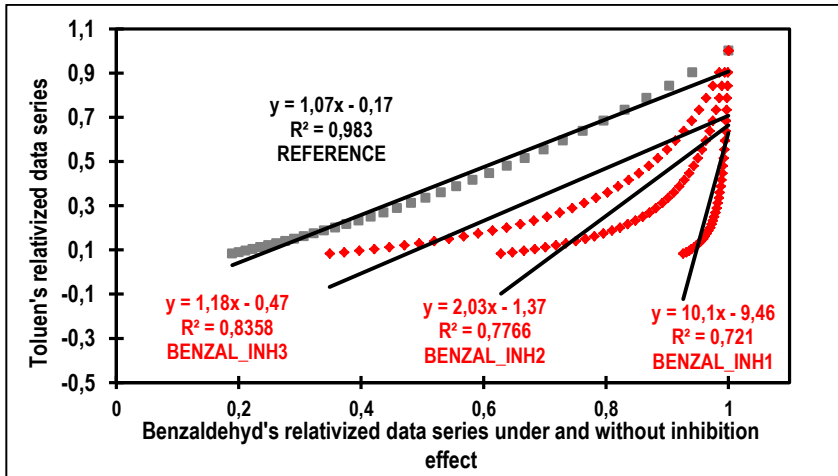


Figure 4.a. Correlations between the normalized concentration data of toluene vs. benzaldehyde. Model data without inhibition (SINGLE) and under various intensity of competitive inhibition effects (BENZAL_INH1-3)

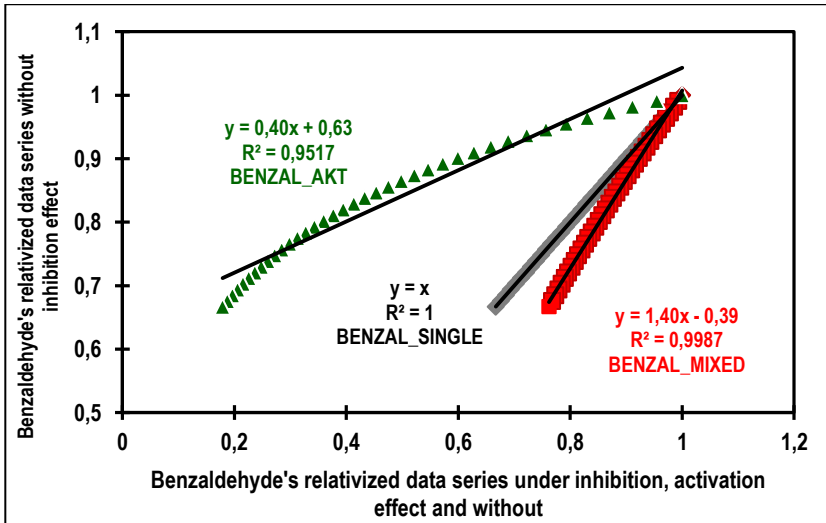


Figure 4.b. Experimental data of independent degradation trials with single substrate degradation (SINGLE) and of experiments with mixed substrates (MIXED) and with activation compound (BENZAL_AKT).

Thesis 4

Using an empirical exponential function between the regression parameters and the concentration of the inhibitor compounds makes it possible to originate the strength of the enzyme activity modification effect.

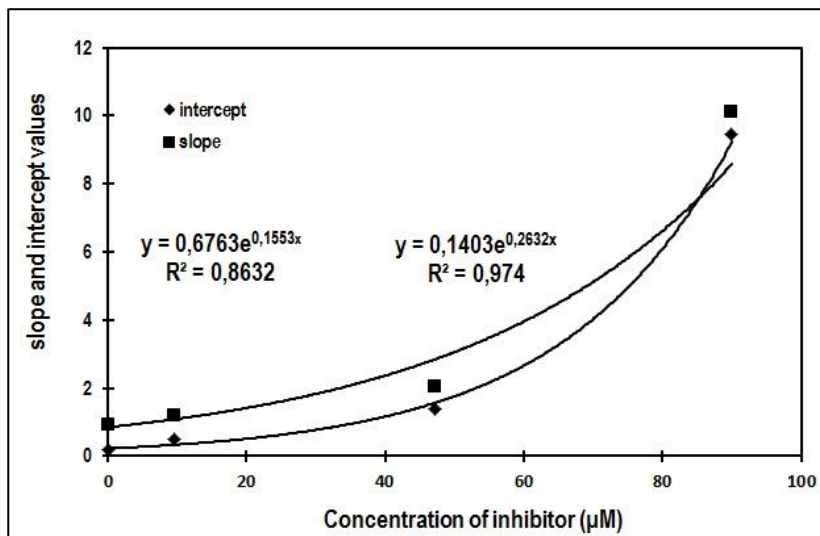


Figure 5. The relationship between the concentrations of the inhibitor compound and the regression parameters, fitting an experimental function

Using this approach further information is available in the field of the exploration of the inhibition effect. Activator- instead of inhibitor concentration is also applicable to measure the strength of the activator effect in the same way.

Thesis 5

Application of PCA method provides an alternative way to indicate the enzyme mechanisms. In consequence of this data representation, the most important information about the kinetic curves appearance on the loadings plot of PC1 and PC2 components.

The processing is applicable to visualize the enzyme activity modification. Using the reference kinetic dataset and its substrate point (without inhibition or activation effect) is always necessary while it is compared to the position of the other substrate points (HERKE et al. 2015).

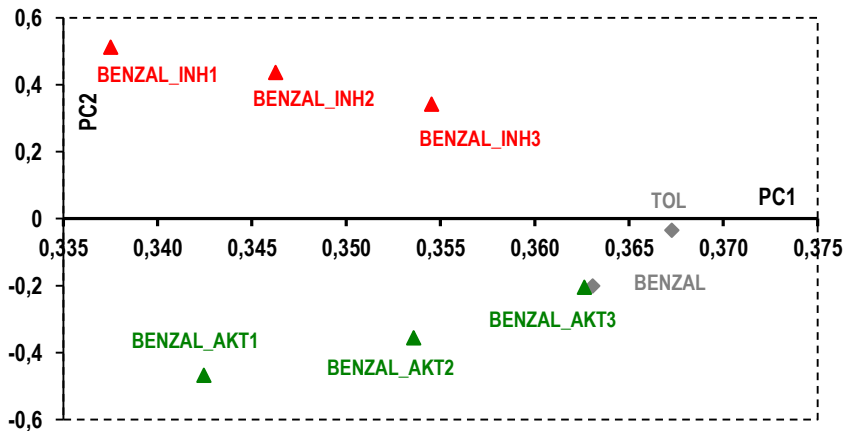


Figure 6. PCA of model biodegradation data. BENZAL and TOL for kinetics without inhibition and BENZAL_INH1-3 for inhibition of different strength, respectively, BENZAL_AKT1-3 for activation of different strength

Related publications of Zoltán Herke

Journal article

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Other publications of Zoltán Herke

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